

1 **Narrow endemics of the Almeria Province (Andalusia, Spain) differ in their traits and**
2 **ecological niche compared to their more widespread congeners**

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20 **Abstract** Understanding the evolutionary ecology of geographically restricted species is of great
21 importance to the design of conservation strategies and to gain insight into the evolutionary
22 processes involved in the origin and maintenance of those species. With a very high proportion
23 of endemic species and/or subspecies, the province of Almeria in Andalusia (SE Spain)
24 represents a prominent hotspot of plant diversity in the Mediterranean Basin. In this paper, we
25 use phylogenetically independent contrasts to examine whether narrow endemics (i.e., species
26 mostly restricted to the Almeria Province, hereafter NE) have evolved a “syndrome of
27 endemism”. Based on published trait values, we test if (i) NE occupy a particular altitudinal
28 range, possess a smaller niche breadth and occur within particular plant associations and if (ii)
29 NE have distinct vegetative and reproductive traits. We find that, compared to their more
30 widespread congeners, NE i) occupy a narrower altitudinal range at the upper or lower limits of
31 their congener’s range; ii) occur in a smaller number of plant communities that are distinct from
32 the communities of widespread species; and iii) have an 11% shorter flowering duration, 28%
33 fewer flowers per inflorescence, 30 % reduced floral display, and 23% lower floral attractiveness.
34 These differences are not explained by a smaller size overall or a different ploidy level. NE have
35 apparently evolved traits that promote selfing. The shift towards higher selfing rates is interpreted
36 as a mechanism to promote reproductive assurance and to decrease gene flow from the more
37 widespread congeners.

38 **Keywords** Endemism • Floral display • Phylogenetically independent contrast • Rarity • Selfing •
39 Traits

40

41 **Introduction**

42

43 Rare species have long fascinated biologists and represent key targets of conservation policies
44 (Kruckeberg and Rabinowitz 1985). Among rare species, endemic species represent the most
45 original component of regional biodiversity, and some endemics, referred to as narrow (or
46 restricted) endemics (NE), have an area of distribution that does not exceed a few square
47 kilometres (Stebbins 1942). Whether rare species possess particular biological attributes has
48 attracted much attention (Gaston and Kunin 1997). Theoretical expectations predict, and a
49 number of empirical studies have confirmed, that NE occupy marginal habitats (Drury 1974;
50 Kruckeberg and Rabinowitz 1985; Hodgson 1986; Verlaque et al. 1997; Lavergne et al. 2003,
51 2004), have adaptations to stressful conditions that in turn result in low competitive ability
52 (Walck et al. 1999), have low dispersal ability (Fiedler 1987; Hanski and Gilpin 1997) and have a
53 breeding system with more selfing (Karron 1987; Kunin and Schmida 1997; Lavergne et al. 2004).
54 Lower genetic diversity in rare compared to common species is one of the most robust
55 generalizations that has emerged from empirical studies (Cole 2003). Possible causes include a
56 higher degree of selfing, low effective population size, demographic bottlenecks and adaptations
57 to narrow ecological conditions (Kruckeberg and Rabinowitz 1985).

58 Comparisons of functional traits between NE and widespread species have generated much
59 data, but generalizations are often difficult to make due to methodological problems. Thus, Beville
60 and Louda (1999) and Murray et al. (2002) in their reviews of 38 and 54 published studies,
61 respectively, were only able to highlight a few robust, general trends. They concluded that
62 patterns of correlation between traits and rarity strongly depend on the phytogeographic and
63 ecological context. Firstly, many published studies are based on only one endemic and one
64 widespread species. Secondly, different studies have examined different traits. Thirdly and most
65 important, studies comparing the endemic flora with the whole flora of the same region suffer
66 from phylogenetic and ecological bias. The endemic flora often has a distinct phylogenetic

67 assemblage and is concentrated in specific habitats (Krukkeberg and Rabinowitz 1985).
68 Therefore, any trait differences between endemics and non-endemics as a whole cannot be
69 unambiguously ascribed to differences in range size.

70 Some regions have an extremely high proportion of endemics in their flora. In the
71 Mediterranean Basin, there are several such “hotspots of endemism” (Médail and Quezel 1999).
72 The flora of southern Spain has one of the highest proportions of endemics in the Mediterranean
73 Basin (553 taxa are endemic to southern Spain, i.e., 15% after Melendo et al. (2003); 480 species
74 are endemic to Andalusia after Mota et al. (2003)). In particular, endemism peaks in the Sierra
75 Nevada, which has the highest concentration of endemic species in Europe (Melendo et al. 2003).
76 Such a high proportion is explained by the high ecological heterogeneity of the region (both
77 climatic and geomorphologic) (Rivas-Martinez et al. 1997, Melendo et al. 2003) combined with its
78 relative geographical isolation since the Pleistocene (Matamala Garcia 2007). In a thorough
79 analysis of the endemic flora of southern Spain, Melendo et al. (2003) showed that the region has
80 a very distinct taxonomic assemblage and is richer in Lamiaceae, Plumbaginaceae, and
81 Brassicaceae but poorer in Poaceae and Apiaceae compared to the general flora. Biological
82 differences with the general flora are striking and include a higher proportion of chamaephytes
83 and a lower proportion of therophytes among the endemics. Entomogamy is overrepresented,
84 and anemogamy is underrepresented in the endemics compared to the non-endemics of southern
85 Spain. Most of these biological differences can be ascribed to differences in the taxonomic
86 assemblage of the endemic flora (Melendo et al. 2003). Therefore, any attempt to analyse the
87 biological traits of the endemic flora of southern Spain must correct for phylogenetic bias.

88 One possible method to overcome such bias is to use a comparative approach based on
89 phylogenetically independent contrasts (Harvey and Pagel 1991). In this approach, a contrast may
90 consist of a pair of related species, one of which is a narrow endemic and the other a widespread
91 species. Thus, the two species in a pair share a broad common genetic background and differ
92 mostly in the size of their range. Any trait differences between them may be related to the

93 contrasting size of their area of distribution. Three studies of the flora of the Mediterranean Basin
94 have used this approach. Kunin and Schmida (1997) investigated 33 pairs of species in the flora
95 of Israel and found that narrow endemics were generally less attractive to pollinators if self-
96 compatible or more attractive if self-incompatible. Lavergne et al. (2003, 2004) examined a broad
97 range of traits in 20 species pairs in Mediterranean France and found that restricted species had
98 attributes characteristic of lower competitive ability, lower reproductive output, and higher selfing
99 rates. They also tended to occupy marginal, rockier habitats. Luna and Moreno (2010)
100 investigated the regeneration niche for 53 species pairs in eastern Spain and found no significant
101 differences between the narrowly distributed and widespread species.

102 In this paper, we use phylogenetically independent contrasts to test if the NE of the flora
103 of eastern Andalusia, more specifically the Almeria Province, have distinct biological and
104 ecological traits compared to their widespread congeners. To that end, we use data from standard
105 floras. Measurements of plant height, leaf size, flower size and number, fruit and seed size, and
106 phenology are used to examine if NE are generally smaller and have distinct reproductive
107 strategies. Lower values of floral display are expected if NE have a breeding system with more
108 selfing (Goodwillie et al. 2010), and seed size and plant height are used as proxies of dispersal
109 capacity (Thompson et al. 2011). Ecological data include altitudinal range (upper and lower
110 altitudinal limits and extension) and the phytosociological units (associations) in which the species
111 occur. These data are used to compare the niches of the NE and the widespread species. In
112 particular, we address the following questions: i) Do NE occupy a narrower niche (i.e., narrower
113 altitudinal extension and fewer phytosociological units)? ii) Do NE occupy a distinct niche; i.e.,
114 do they occur in distinct plant communities not occupied by widespread species, and do they
115 occupy a distinct altitudinal range?

116

117 **Materials and methods**

118

119 *Phytogeographic context*

120

121 The province of Almeria is geographically delimited in the E and the S by the Mediterranean Sea,
122 in the W and NW by mountain chains oriented W-E, the Sierra Nevada, Sierra de Gador, Sierra
123 de Los Filabres, and in the N by the Sierra de las Estancias and Sierra de María. The province is
124 extremely diversified in terms of climate due to its broad altitudinal range. Its coastal region
125 represents the most arid area in Europe with a mean annual rainfall of 289 mm and a mean
126 annual temperature of 18.1°C in Almeria. In contrast, mean annual temperature is 3.8°C, and
127 mean annual rainfall is 705 mm in the upper slopes of the Sierra Nevada. The area of this study
128 belongs to three biogeographic sectors with different percentages of endemism: Nevadense
129 (15.5%), Alpujarreño-Gadoreño (2.5%) and Almeriense (6.6%) (Rivas-Martinez et al. 1997).

130 A check-list of species endemic to the Almeria Province has been established based on the
131 Flora Iberica (Castroviejo et al. 1986-2012) and the digital Flora Vasculare de Andalucía Oriental
132 (Blanca et al. 2011). However, as basing endemism on administrative boundaries makes little
133 sense, the list has been extended to species whose distribution extends to the Sierras in the
134 adjoining provinces of Murcia or Granada. Widespread, congeneric species have been selected
135 from the flora of the Almeria Province as follows. A widespread species is defined here as a
136 species whose distribution area covers at least five provinces; a more restrictive criterion (i.e., a
137 wider distribution range in the widespread species) would have decreased the number of pairs.
138 Another restrictive ecological criterion has been applied; the altitudinal range of the widespread
139 and the NE species must at least partly overlap, which ensures that the congeners do not occur in
140 completely different ecogeographic contexts. The latter criterion excluded a few narrow endemics
141 restricted to the upper slopes of the Sierras whose traits reflect the specific environmental filter
142 of those alpine conditions (Penas et al. 2005). Distributional data were obtained from the

143 aforementioned floras and supplemented by the Anthos database (Aedo and Castroviejo 2009)
144 and SIVIM (www.sivim.info/sivi/).

145 For five genera, published phylogenies were used to select the most closely related
146 widespread species (*Arenaria*: Valcàrcel et al. 2006; *Centaurea*: Víctor et al. 2007; *Genista*: Pardo et
147 al. 2004; *Festuca*: Torrecilla and Catalàn 2002; *Carex*: Escudero et al. 2008). For the other genera,
148 no phylogeny is available, or the published phylogenies do not include our endemic species. In
149 such cases, the widespread species were based on taxonomic relatedness (same section or
150 subsection in the genus) following Kelly et al. (1996). Finally, the congeners have been selected to
151 have the same life form, pollination mode and dispersal mechanism. These restrictive criteria aim
152 at maximizing the genetic relatedness of the congeners and reducing confounding sources of
153 variation not directly related to the extent of distribution area (Baskauf et al. 2001). By using
154 restrictive selection criteria, we deliberately focus on the more subtle ecological differences
155 between NE and widespread species compared to the broad differences in the life form spectrum
156 highlighted by Melendo et al. (2003).

157 A total of 25 species pairs fulfilling all of the selection criteria have been found. Two pairs have
158 been selected from the genus *Teucrium*, but they belong in different sections. For two genera,
159 *Verbascum* and *Helianthemum*, two widespread species were found to fulfil our criteria. In those
160 two cases, the traits of the NE were compared to the average value of the traits of the two
161 widespread congeners.

162 Due to the restrictive selection criteria, our sample might be biased towards neoendemic
163 species, i.e., recently diverged species that have close relatives in the regional flora, but
164 paleoendemic species are not a prominent component of the flora of southern Spain (Médail and
165 Diadema 2009). Most major clades of flowering plants are represented in our sample. The
166 overrepresentation of Lamiids (8 of 25 pairs) in our sample reflects the particular taxonomic
167 assemblage of the endemic flora of Andalusia (Melendo et al. 2003).

168

169 ***Traits***

170

171 Species' trait values were obtained from two recent standard floras (Castroviejo et al. 1986-2012;
172 Blanca et al. 2010). No data were available for our species in the TRY database (Kattge et al.
173 2011). We aimed to sample a broad range of traits, describing the vegetative parts (leaf size, plant
174 height) and reproductive attributes (flower size and number, seed size). All of the traits that were
175 available for at least 10 species pairs have been included in our database.

176 Flora trait values are mostly reported as min-max ranges and are sometimes supplemented
177 by extreme values (e.g., height: (15-)25-35(-45) cm). All analyses have been conducted for the
178 minimum value, the maximum value, the arithmetic mean of the minimum and the maximum
179 values and the arithmetic mean of the extreme values. As the results for the different values were
180 essentially the same, only those for the arithmetic mean of the min-max range are reported.

181 Vegetative traits include plant height, leaf length (L), leaf width (w) and the L*w product as
182 a proxy of leaf area. Reproductive traits include the number of flowers per inflorescence,
183 inflorescence length, corolla and calyx length, fruit length and width, and seed length. The floral
184 display was calculated as the product of the number of flowers per inflorescence * corolla length
185 (Goodwillie et al. 2010) (petal length was substituted for corolla length for *Alyssum*, *Arenaria*,
186 *Brassica*, *Helianthemum*, *Potentilla*, *Seseli*, and *Silene*; corolla diameter for *Dianthus*, *Limonium*, and
187 *Ranunculus*; upper petal length for *Genista* and *Ulex*). The floras describe inflorescence size in
188 terms of either length or number of flowers, so it is not possible to obtain both inflorescence
189 length and floral display for all of the species in our sample. However, these traits capture similar
190 information (Schoen et Dubuc (1990); Goodwillie et al. (2010)). Therefore, inflorescence length
191 and floral display have been pooled into a combined variable referred to as "attractiveness" to
192 increase sample size. The duration of flowering was calculated as the number of months between
193 the earliest and the latest month of flowering as reported in the aforementioned floras. Ploidy
194 level was also obtained from the floras and from Escudero et al. (2008) for *Carex*.

195

196 *Ecological niche*

197

198 Assuming that syntaxonomic diversity is an indicator of ecological diversity, we used
199 phytosociological data as surrogates for niche description (Redzic 2007). The data were obtained
200 from the SIVIM database (www.sivim.info/sivi/). The number of phytosociological units (plant
201 associations) in which a species is recorded in the province of Almeria was used as a proxy for
202 niche breadth. However, that number is biased by sampling effort. To correct for sampling bias,
203 we used a resampling method without replacement. For each species pair, if the NE is recorded
204 in r_{NE} relevés representing a total of a_{NE} associations and the widespread species is recorded in r_W
205 relevés representing a_W associations (with $r_{NE} < r_W$), the r_{NE} relevés with the widespread species
206 were randomly resampled 100 times, and the average of the number of associations was used as
207 the unbiased value. Sample size was as follows: *Alyssum* (9), *Arenaria* (104), *Carex* (70), *Centaurea*
208 (6), *Coris* (15), *Dianthus* (9), *Genista* (38), *Helianthemum* (8), *Herniaria* (65), *Potentilla* (16), *Ranunculus*
209 (13), *Sideritis* (9), *Teucrium* (21), and *Ulex* (13).

210 The altitudinal niche was also analysed, using the upper and lower altitudinal limits (in m
211 above sea level) from the abovementioned floras. Altitudinal range, a descriptor of niche breadth
212 (number of meters of extension), and mean altitude, a descriptor of niche location (arithmetic
213 mean of upper and lower altitudinal limit), have been calculated for each species.

214

215 *Data analysis*

216 To test the null hypothesis that NE species and their congeners have the same trait values, paired
217 Student t-tests were performed. When differences between species were not normally distributed
218 (Shapiro test), a paired Wilcoxon signed rank test was applied.

219 For chromosome number, the data could not be analysed in the same way because some
220 species had two different numbers. In such cases, the different cytotypes of the same species

221 were treated as different taxa, and new pairs were created accordingly. As this resulted in
222 pseudoreplication, a conservative test was applied, and the number of degrees of freedom in the
223 Wilcoxon test was not increased.

224 Because the analysis of individual traits resulted in many simultaneous tests and did not
225 consider correlations among traits, a multivariate analysis was also performed. Principal
226 component analysis was conducted on the correlation matrix of a subset of the 7 variables and 10
227 species pairs without missing values (corolla length, floral display, flowering duration, flower
228 number, leaf area, fruit length, and height; genera included: *Arenaria*, *Brassica*, *Chaenorbinum*,
229 *Helianthemum*, *Herniaria*, *Linaria*, *Potentilla*, *Ranunculus*, *Seseli*, and *Verbascum*). Ecological data were
230 not included in this analysis. The first and second principal components were then used as new
231 synthetic variables and subjected to paired t-tests. All statistical analyses were performed with R
232 (version 2.14.0) (R Development Core Team 2009).

233

234 **Results**

235

236 Of the 18 variables considered, 7 showed a statistically significant difference between NE and the
237 widespread species (i.e., altitudinal range, lower altitudinal limit, number of communities,
238 flowering duration, number of flowers, floral display, and attractiveness) (Table 2, Fig. 1).

239

240 ***Ecological traits***

241 On average, NE had a 52% narrower altitudinal range (611 m) compared to widespread species
242 (1515 m) (Table 2, Fig. 1a), and this difference was true for 22 of 25 species pairs. The altitudinal
243 distribution of NE and widespread species were strikingly different (Fig. 2). Thus, NE showed a
244 strongly bimodal distribution frequency as they were concentrated at both ends of the altitudinal
245 gradient (0-450 m and 2400-2900 m, respectively) (Fig 2a). Widespread species had a unimodal
246 distribution that peaked at intermediate altitude (900-1300 m) (Fig. 2b).

247 The number of phytosociological units was significantly lower in NE compared to
248 widespread species, ranging from 2 to 24 in widespread species and from 1 to 13 in NE (Table 2,
249 Fig. 1b). The number of units was equal or smaller in the NE in 14 of 15 pairs with *Herniaria*
250 being the only exception. In seven pairs, the endemic and the widespread species did not occur in
251 the same association. In seven other pairs, only one association was shared between the two
252 species, and in the case of *Thymus*, four associations were shared (Fig. 3).

253
254 ***Reproductive traits***

255 Flowering duration was 11% (i.e., ca. one month) shorter in NE (Table 2); the opposite trend was
256 observed in only five pairs (Fig. 1c). The number of flowers per inflorescence was 28% lower in
257 NE (Table 2) (i.e., lower in 7 pairs, equal in 2 pairs and higher in 1 pair (*Potentilla*)) (Fig. 1d).
258 Floral display was 29% smaller in NE (Table 2) (Fig. 1e). NE had a higher floral display value in
259 only two pairs, namely, *Ranunculus* and *Chaenorhinum*. For the combined “attractiveness” variable,
260 NE had 23% lower values (Table 2, Fig. 1f). NE had higher values of this trait in only three pairs
261 (*Potentilla*, *Ranunculus*, and *Chaenorhinum*). Inflorescence length was lower in the NE in four cases,
262 equal in three cases and higher in only one of eight cases (*Teucrium* sect. *Simplicipilosa*), but the
263 difference (10.5 cm versus 8.7 cm) was not significant (Table 2). Corolla length, calyx length, fruit
264 length and width, and seed size were not significantly different between NE and widespread
265 species (Table 2).

266

267 ***Vegetative traits***

268 No statistically significant difference was found between NE and widespread species for
269 vegetative traits (leaf length and width, leaf area, and plant height) (Table 2).

270

271 ***Ploidy level***

272 Chromosome numbers are available for 20 pairs (Table 1). In 12 pairs, the two species had the
273 same ploidy level. In seven pairs, the widespread species had several ploidy levels (two or three),
274 but this was found in only 1 NE, and the difference in the proportions (i.e., 1 of 20 vs. 7 of 20)
275 was significant (Fisher's exact test, two-sided: $P=0.048$). In the pairwise comparison tests, the
276 different cytotypes of the same species were treated as different taxa. For the 30 new pairs, the
277 results of the Wilcoxon signed-rank test were not significant ($V=58$, $P=0.75$ $n=20$).

278

279 ***Principal Component Analysis***

280 PC1 and PC2 accounted for 29% and 25% of the variance, respectively. PC1 was mostly
281 correlated to leaf area, fruit length and corolla size (Fig. 4a), and it captures most of the
282 phylogenetic source of variation. Thus, genera with large flowers, large leaves and large fruits
283 (e.g., *Verbascum*) are mostly located on the left side of the scatter plot while genera with the
284 opposite combination of traits (e.g., *Herniaria*) are on the right side of PC1 (Fig. 2a). PC2 was
285 negatively correlated to floral display, flowering duration and the number of flowers per
286 inflorescence. Compared to their widespread congeners, NE had significantly higher scores on
287 PC2 (paired t-test: $t = 3.17$, $df = 9$, $P = 0.011$), and this was true in eight of ten pairs. In one pair,
288 there was no tendency (*Potentilla*), and in one pair, the difference was in the opposite direction
289 (*Ranunculus*). NE also had a higher score on PC1 (verified in 7 of 10 pairs), but the difference was
290 not significant (paired t-test: $t = 1.92$, $df = 9$, $P = 0.086$).

291

292 **Discussion**

293

294 **Narrow endemics occupy narrower, marginal niches**

295

296 We found that endemics have narrower altitudinal ranges than their widespread congeners. This
297 result fits well with the general correlation between niche breadth and the extent of geographical
298 range in plants (Slatyer et al. 2013), which was verified for endemic species in the Alps (Essl et al.
299 2009). That pattern is demonstrated here for the first time in a multispecies, phylogenetically
300 controlled comparison. Much more interestingly, the altitudinal niche of endemics is not
301 randomly nested within the niches of their widespread congeners. Endemics are strikingly more
302 frequent at both ends of the elevational gradient in sharp contrast with the distribution of
303 widespread species, which peak at mid-altitudes. The positive correlation between altitude and
304 the proportion of endemics in the flora of the Sierra Nevada has recently been confirmed by
305 Fernandez-Calzado et al. (2014). The low altitude endemics in our study are narrow endemics of
306 the Almeriense sector (Rivas-Martinez et al. 1997). Compared to the regional climate of
307 Andalusia, these two groups of endemics occur in extreme conditions with cooler temperatures
308 and a shorter growing season at high altitude (oro-Mediterranean humid climate) and hot,
309 extremely dry conditions in the region of Cabo de Gata (thermo-Mediterranean semiarid climate)
310 (Rivas-Martinez et al. 1997). Thus, with respect to altitude and climate, narrow endemics in the
311 Almeria Province occupy marginal niches compared to their widespread congeners.

312 Niches were also assessed by means of phytosociological units. The results are again clear-cut
313 with endemics occurring in 28% fewer phytosociological units. Furthermore, they also tend to
314 occur in distinct phytosociological units. Assuming that phytosociological diversity indicates
315 ecological diversity (Redzic 2007), our results suggest that endemics are indeed ecologically
316 specialised species. Taken together, the altitudinal and phytosociological distributions of
317 endemics in this study are consistent with earlier generalisations that endemics occupy marginal

318 and/or stressful habitats (Drury 1974; Kruckeberg and Rabinowitz 1985; Lavergne et al. 2003,
319 2004; Medail & Verlaque 1997), including Baetic high mountains (Penas et al. 2005). In the flora
320 of California, the ranges of narrow endemic species are often nested within those of their more
321 widespread sister species. This suggests that narrow endemics have sympatrically diverged in a
322 process referred to as “budding speciation” (Anacker & Strauss 2014). In the Almeria Province,
323 however, the situation is somewhat different as the endemics are markedly concentrated at the
324 two extremes of an altitudinal gradient. This may indicate that narrow endemics have originated
325 by ecological speciation in peripheral populations subjected to divergent natural selection
326 (Schluter 2009).

327

328 Adaptation to extreme habitats might entail a cost in terms of decreased competitive ability in the
329 narrow endemics (Medail & Verlaque 1997; Imbert et al. 2011). However, we found no clear
330 evidence for reduced vegetative vigour or competitive ability in endemics as they did not have
331 smaller heights or smaller leaves. This contrasts with the French Mediterranean region where
332 endemics occupy marginal habitats (rocky slopes with sparse plant cover) and do have reduced
333 competitive ability compared to their widespread counterparts (Lavergne et al. 2003, 2004).
334 However, widespread species in southern France occur in much more mesic habitats, in deeper
335 soil and in dense vegetation cover. Such habitats are scarce in the Almeria Province, so the
336 contrast in ecological conditions between NE and widespread species is less striking with both
337 endemic and widespread species occurring in arid habitats with shallow soils in open,
338 unproductive communities. Endemic species may also have a restricted geographic range due to
339 low dispersal ability (Youssef et al. 2011). We found no significant difference in propagule size or
340 plant height, two predictors of dispersal capacity for species sharing the same dispersal
341 mechanisms (Fenner & Thompson 2005; Thompson et al. 2011). Therefore, lower dispersal
342 capacity is less likely to account for the much more restricted geographic range of endemics
343 compared to their congeners.

344 However, narrow endemics might still have lower reproductive output (Murray et al. 2002;
345 Lavergne et al. 2004) resulting in low propagule pressure and lower colonizing ability, but field
346 measurements are needed to test this hypothesis.

347

348 *Are endemics more prone to selfing?*

349

350 Lower attractiveness to pollinators in narrow endemics is a striking result of this study.
351 Interestingly, smaller floral displays and inflorescence size are not by-products of smaller plant
352 size because plant height, leaf size and calyx size did not differ between endemics and widespread
353 species. Our results indeed suggest a breeding system with more selfing in narrow endemics.
354 Species existing as small, isolated populations or those adapted to specific habitats are more
355 prone to evolving a selfing breeding system than species with the opposite attributes (Kunin and
356 Schmida 1997; Karron 1987; Lavergne et al. 2003, 2004; Lowry and Lester 2004). Other studies
357 have also found lower attractiveness to pollinators in rare species (Kunin and Schmida 1996;
358 Lavergne et al. 2004; Rymer et al. 2005), and these findings suggest that lower attractiveness to
359 pollinators in the endemics has evolved in response to selective pressure (Charlesworth and
360 Charlesworth 1981; Goodwillie et al. 2010). The shorter flowering duration in the endemics in
361 this study is also consistent with a breeding system with more selfing (Bolmgren et al. 2003;
362 Elzinga et al. 2007; Rathcke and Lacey 1985). Shorter flowering duration in rare species was also
363 found by Murray et al. (1999), Lahti et al. (1991), Cadotte and Lovett-Doust (2002) and
364 Gabrielova et al. (2013). The elevation of self-fertilization in narrow endemics may facilitate their
365 invasion of marginal habitats by reducing the level of maladaptive gene flow from nearby
366 populations and by conferring reproductive assurance (Levin 2010). Transition to selfing may
367 itself be a driver of speciation (Wright et al. 2013), and our results are consistent with a scenario
368 of selfing endemics evolving from widespread outcrossers.

369 Pollen limitation is another selective force promoting evolution towards increased rates of
370 selfing (Knighth et al. 2005). Biodiversity hotspots are known to be prone to pollen limitation
371 (Vamosi 2006), and Alonso et al. (2010) demonstrated that transitions to endemism in hotspots
372 are generally associated with transitions to self-compatibility. Endemic species may thus be less
373 attractive to pollinators (Knighth et al. 2005) and particularly vulnerable to pollen limitation due to
374 their low population density (Brown 1984; Gaston et al. 2000), marginal habitats, or short
375 flowering duration (Cadotte and Lovett-Doust 2002 Gabrielova et al. 2013, Lahti et al. 1991,
376 Murray et al. 1999). An outcrossing breeding system may increase the risk of receiving
377 heterospecific pollen, which may be critical in endemics that exist as small populations in contact
378 with close relatives; this is indeed true in the Almeria Province (Harder et al. 1993). Therefore,
379 for autocompatible species, increased selfing is expected to evolve in narrow endemics that co-
380 occur with close relatives in biodiversity hotspots. In contrast, self-incompatible endemic species
381 cannot escape pollen limitation by selfing, so they actually evolve towards more attractive floral
382 traits (Kunin and Schmid, 1996). In our study, attractiveness was higher in the endemic in only
383 three pairs, i.e., *Ranunculus*, *Teucrium* and *Chaenorbinum*. Interestingly, *Ranunculus* is known to be
384 self-incompatible (Lundquist 1990).

385

386 **Conclusions**

387

388 In conclusion, our comparative approach based on phylogenetically independent contrasts has
389 revealed few differences between endemics and their widespread relatives, but the differences in
390 ecology and traits are mutually consistent. Endemics of the Almeria Province appear to be
391 narrow-niched ecological specialists, and they exhibit smaller floral displays that are most likely
392 related to increased selfing. Altogether, these results are consistent with a scenario of
393 neoendemism resulting from the parapatric divergence of populations at the ecogeographical

394 margins of the range of widespread species. Field population biology and phylogeographic
395 studies are needed to test this hypothesis.

396

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402

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Table 1 The 25 studied species pairs (the first species in a pair is the endemic species). Nomenclature follows Castroviejo *et al.* (1986-2012) and Blanca *et al.* (2009). Number of populations, number of individuals and occupancy area are from Bañares *et al.* (2010), Blanca *et al.* (1999, 2000), Cabezudo *et al.* (2005), Mota *et al.* (2003), and Moreno (2008). Number of chromosomes is from Castroviejo *et al.* (1986-2012), Blanca *et al.* (2009) and for the genus *Carex*, Escudero *et al.* (2008).

Life form: Na=nanophanerophyte, Ch=chamaephyte, He=hemicryptophyte, Ge=geophyte, Th=therophyte

Dispersal: EX=exozoochorous, MY=myrmecochorous, Ane=anemochorous, Bar=Barochorous

Pollination: Ent=Entomogamous, Ang=Anemogamous. Reproductive system: hermaphrodite except if “+ g”: gynodioecious

Geographic range: Al=Almeria Province, Gr=Granada Province, sector Nevadense, Mu=Murcia Province

Family	Genus	Species	N. pop.	N. indiv.	Occupancy area (km ²)	N. Chrom.	Geographic range	Life form/dispersal/pollination
Apiaceae	<i>Seseli</i>	<i>S. intricatum</i> Boiss. <i>S. montanum</i> subsp. <i>montanum</i> Willk.	5	5000	195	22 22, 33	Al W Medit Basin	He/Bar/Ent, Ane He/Bar/Ent
Asteraceae	<i>Centaurea</i>	<i>C. sagredoii</i> Blanca <i>C. monticola</i> Boiss. ex DC.	2	149	65	18 18	Al SE Spain and Morocco	He/MY/Ent He/MY/Ent
Brassicaceae	<i>Abyssum</i>	<i>A. gadorense</i> P.Küpfert <i>A. montanum</i> L.	5	124	20	8, 18, 24	Al W Medit Basin	Ch/Bar/Ent Ch/Ane/Ent
Brassicaceae	<i>Brassica</i>	<i>B. repanda</i> subsp. <i>almerienses</i> Gomez Campo <i>B. repanda</i> subsp. <i>blancoana</i> (Boiss.) Heywood	3	331	12	20 20	Al. (Gr) E, S & SE Spain	He/Bar/Ent He/Bar/Ent
Caryophyllaceae	<i>Arenaria</i>	<i>A. tetraquetra</i> subsp. <i>amabilis</i> (Bory) H. Lindb <i>A. armerina</i> subsp. <i>armerina</i> Bory				40 28, 30	Al. (Gr) SE Spain	Ch/Bar/Ent Ch/Bar/Ent
Caryophyllaceae	<i>Dianthus</i>	<i>D. charidemi</i> Pau. <i>D. broteroi</i> Boiss. et Reut.	17	5000	1229	60	Al. (Mu) S Spain	Ch/Bar/Ent Ch/Bar/Ent
Caryophyllaceae	<i>Herniaria</i>	<i>H. boissieri</i> J. Gay <i>H. scabrida</i> Boiss.	5	750	2.8	18	Al. (Gr) Spain, Portugal	Ch/Bar/Ent Ch/Bar/Ent
Caryophyllaceae	<i>Silene</i>	<i>S. littorea</i> subsp. <i>adscendens</i> (Lag.) Rivas Goday <i>S. littorea</i> subsp. <i>littorea</i> Brot.	22	3740	7147	24	Al Spain, Portugal, N Afr.	Th/Bar/Ent Th/Bar/Ent
Cistaceae	<i>Helianthemum</i>	<i>H. alypoides</i> Lisa et Rivas Goday <i>H. birtum</i> Mill. <i>H. violaceum</i> Lange	7	2000000	2045	20 20	Al W Medit Basin W Europe	Ch/Bar/Ent Ch/EX/Ent Ch/EX/Ent
Cyperaceae	<i>Carex</i>	<i>C. camposii</i> Boiss et Reut. <i>C. mairei</i> Coss. et Germ.	3	750	9	72 68, 69, 70	Al E Spain	He/Bar/Ang He/Bar/Ang
Fabaceae	<i>Genista</i>	<i>G. versicolor</i> Boiss. <i>G. longipes</i> subsp. <i>longipes</i> Pau.				36 18	Al et Gr SE Spain	Na-Ch/Bar/Ent Ch/Bar/Ent

Fabaceae	<i>Ulex</i>	<i>U. canescens</i> Lange <i>U. parviflorus</i> subsp. <i>parviflorus</i> Pourr.	9	3241	1150	32 32	Al Andalusia	Na/MY/Ent Na/MY/Ent
Lamiaceae	<i>Nepeta</i>	<i>N. nepetella</i> subsp. <i>laciniata</i> <i>N. nepetella</i> subsp. <i>murcica</i>				34	Al. (Gr) SE Spain and Morocco	Ch/Bar/Ent Ch/Bar/Ent
Lamiaceae	<i>Sideritis</i>	<i>S. hirsuta</i> L. <i>S. osteoxylla</i> (Pau.) Alcaraz et al.				26 28, 56	W Medit Basin Al	Ch/Ane/Ent Ch/Ane/Ent
Lamiaceae	<i>Teucrium</i> sect. <i>Polium</i>	<i>T. hieronymi</i> Sennen <i>T. capitatum</i> L.				39 26, 39, 52	Al Spain, Portugal	Ch/EX/Ent + g Ch/EX/Ent + g
Lamiaceae	<i>Teucrium</i> sect. <i>Simplicipilosa</i>	<i>T. charidemi</i> Sandwith. <i>T. eriocephalum</i> Willk. subsp. <i>eriocephalum</i>	16	2180	4591	26 26	Al SE Spain	Ch/EX/Ent Ch/EX/Ent
Lamiaceae	<i>Thymus</i>	<i>Th. hyemalis</i> subsp. <i>millefloris</i> (Rivera et al.) R. Morales <i>Th. baeticus</i> Boiss. ex Lacaita	2	828	1000	58 58	Al S & SE Spain	Ch/Ane/Ent Ch/Ane/Ent
Plantaginaceae	<i>Chaenorbinum</i>	<i>C. grandiflorum</i> (Coss.) Willk. subsp. <i>grandiflorum</i> <i>C. minus</i> subsp. <i>minus</i> (L.) Lange	9	-	1375	14	Al Cosmop.	Th/-/Ent Th/Ane/Ent
Plantaginaceae	<i>Linaria</i>	<i>L. oblongifolia</i> subsp. <i>benitoi</i> (Fern. Casas) L.Sáez, M.B.Crespo & M.Bernal <i>L. oblongifolia</i> Boiss et Reut. subsp. <i>haenseleri</i>	4	4000	1000	12 12	Al S Spain & Portugal	Th/-/Ent Th/-/Ent
Plumbaginaceae	<i>Limonium</i>	<i>L. tabernense</i> Erben <i>L. cossonianum</i> Kuntze	10	20150	5006	16 16	Al SE Spain + Balearic isl.	Ch/Ane/Ent Ch-He/Ane/Ent
Poaceae	<i>Festuca</i>	<i>F. pseudoeskia</i> Boiss. <i>F. scariosa</i> (Lag.) Pau				14 14	Al. (Gr) SE Spain	He/Ane/Ang He/Ane/Ang
Primulaceae	<i>Coris</i>	<i>C. hispanica</i> Lange <i>C. monspeliensis</i> L.	12	5000	1088	18 18, 56	Al W Medit Basin	Ch/Ex/Ent Ch/Bar/Ent
Ranunculaceae	<i>Ranunculus</i>	<i>R. acetosellifolius</i> Boiss. <i>R. demissus</i> DC.	7	1950	87	16 16	Al. (+Gr) Cosmop. (mountains)	Ge/EX/Ent Ge/EX/Ent
Rosaceae	<i>Potentilla</i>	<i>P. nevadensis</i> Boiss. <i>P. birta</i> L.	8	387	14	14, 28 14	Al. (Gr) W Medit Basin	He/Bar/Ent He/Bar/Ent
Scrophulariaceae	<i>Verbascum</i>	<i>V. charidemi</i> Murb. <i>V. simplex</i> Labill. <i>V. virgatum</i> Stokes	17	223	30	32 62, 64, 66	Al W Medit Basin Cosmop.	He/Bar/Ent He/-/Ent He/EX/Ent

1 **Table 2** Comparison of traits and ecological ranges between narrow endemic (NE) and
2 widespread (W) species in Almeria Province (SE Spain): phylogenetically independent contrasts.
3 n=number of pairs. The magnitude of difference is calculated as the difference between the value
4 of the endemic and the widespread species, expressed in percentage. The results are from
5 Student's paired t-test or Wilcoxon matched-pairs signed-rank test if the data were not normally
6 distributed.
7

	Student test t=	Wilcoxon test W=	n	P	NE vs. W	Magnitude of difference
Ecological range						
Altitudinal range	-3.17		25	$6 \cdot 10^{-7}$	E < W	52%
Upper altitudinal limit	-1.77		25	0.09		
Lower altitudinal limit		17.5	25	0.00042	E > W	57%
Number of plant communities	4.17		17	0.0034	E < W	32%
Reproductive traits						
Flowering duration	2.20		24	0.038	E < W	11%
Number of flowers per inflorescence	2.71		10	0.024	E < W	28%
Inflorescence length	1.49		8	0.19		
Calyx length		57	20	0.59		
Corolla length		130	23	0.63		
Floral display	3.25		13	0.0069	E < W	29%
Attractiveness		156	21	0.0023	E < W	23%
Fruit length		123	22	0.74		
Fruit width	1.11		14	0.28		
Seed length	0.79		11	0.48		
Vegetative traits						
Height	1.58		25	0.47		
Leaf length		162	23	0.25		
Leaf width		123	23	0.81		
Leaf length x Leaf width	-0.02		22	0.99		

9 Figures

10

11 **Fig. 1** Comparison of ecological attributes and traits of endemic species of the province of
12 Almeria and widespread congeners. Each point represents a species pair. The results of paired
13 comparison tests (Student t-test or Wilcoxon test) are indicated. The first bisector represents
14 equal trait values for endemic and widespread species.

15

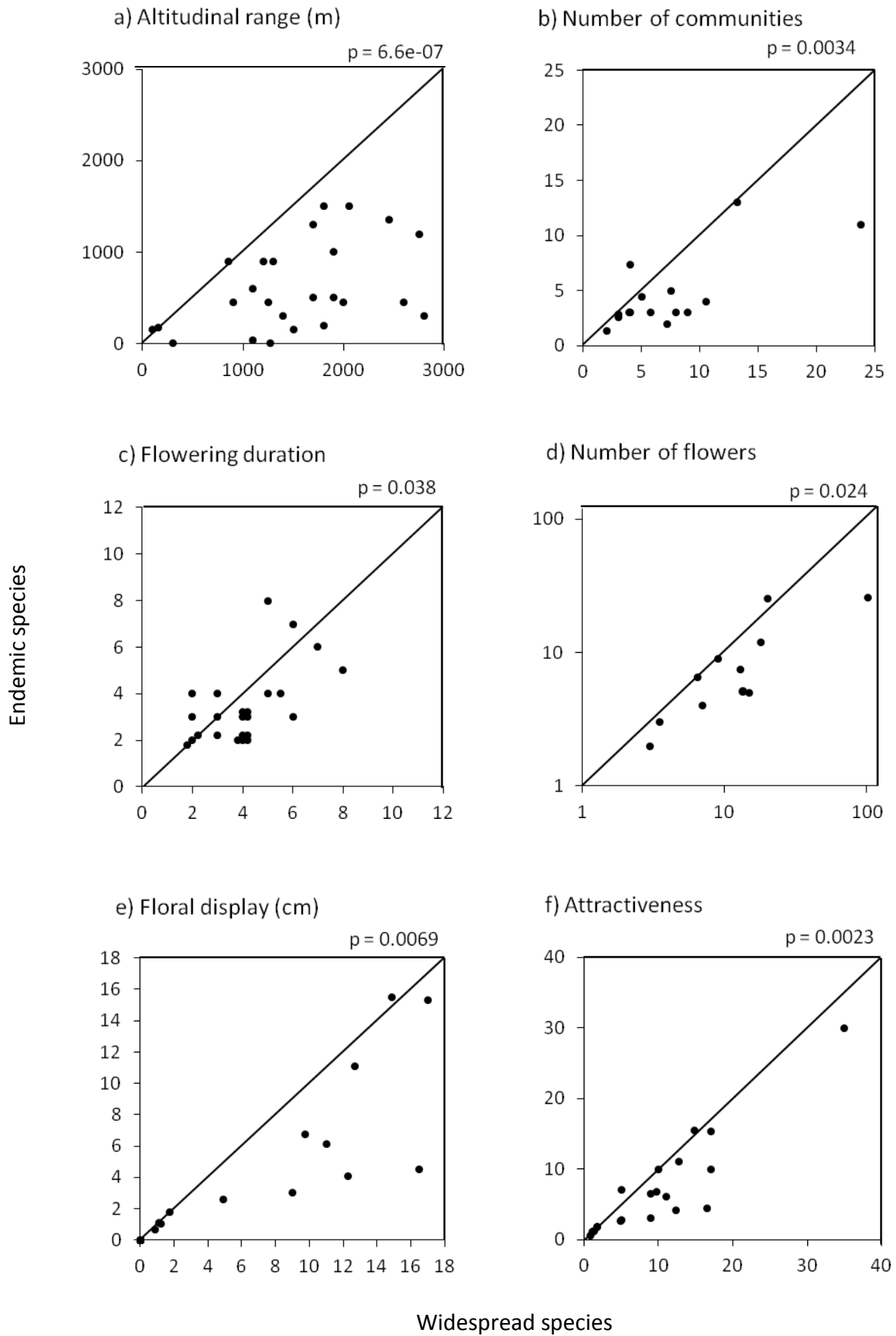
16 **Fig. 2** Frequency distribution of the mean altitude of the occurrence of narrow endemic species
17 and widespread congeners in the province of Almeria.

18

19 **Fig. 3** Number of plant associations in which narrow endemic species and widespread congeners
20 occur in the province of Almeria and number of shared associations (corrected for differences in
21 sample size).

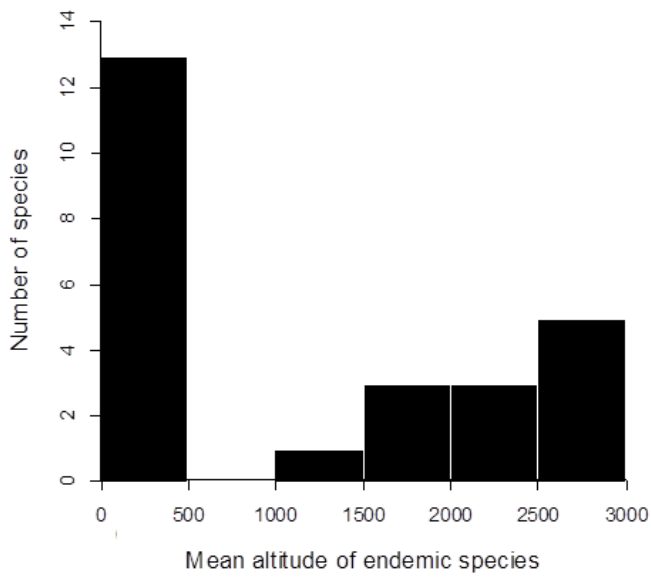
22

23 **Fig. 4** Principal component analysis of seven traits in ten species pairs consisting of an endemic
24 species of the province of Almeria and a widespread congener. a) Scatterplot of species; the two
25 congeners are linked by a vector (arrow: widespread to endemic). b) Trait loadings on PC1 and
26 PC2; the red vector represents the average of the 10 species pairs. Traits included: Corolla
27 (corolla length); leaf area (leaf length x leaf width), Fruit (fruit length), Height (vegetative height),
28 Flo_disp (floral display), Flo_dur (flowering duration), Flo_nb (flower number); Genera included
29 (widespread-endemic): *Seseli* (1-2), *Herniaria* (3-4), *Arenaria* (5-6), *Ranunculus* (7-8), *Potentilla* (9-10),
30 *Chaenorbinum* (11-12), *Verbascum* (13-14), *Linaria* (15-16), *Helianthemum* (17-18), *Brassica* (19-20).

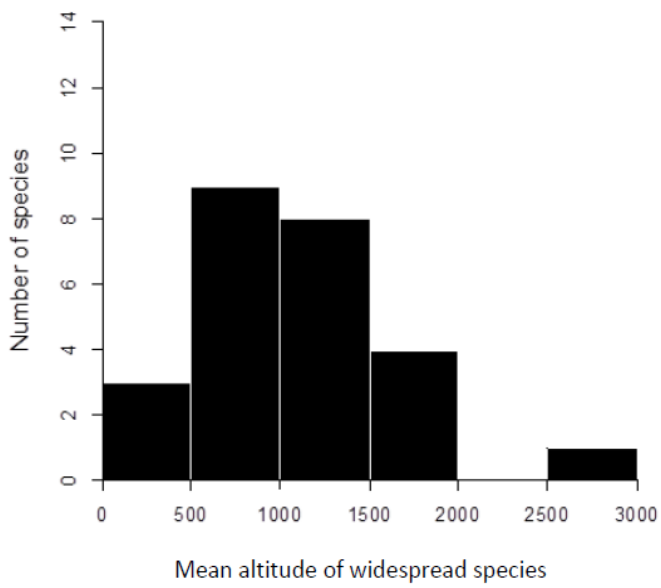


32 Figure 2

33

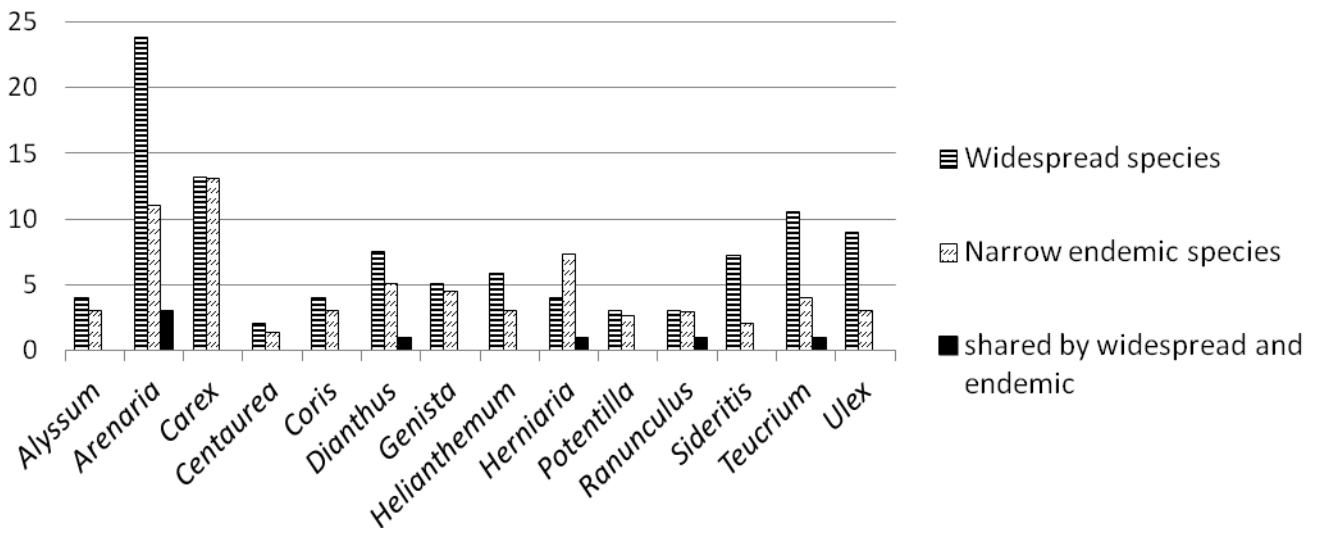


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35 Figure 3

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